

1 **Joint Retrieval of PM<sub>2.5</sub> Concentration and Aerosol Optical Depth over China**  
2 **Using Multi-Task Learning on FY-4A AGRI**

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## ABSTRACT

17  
18 Aerosol optical depth (AOD) and fine particulate matter with a diameter of less than  
19  $2.5\mu\text{m}$  ( $\text{PM}_{2.5}$ ) play crucial roles in air quality, human health, and climate change.  
20 However, the complex correlation of AOD– $\text{PM}_{2.5}$  and the limitations of existing  
21 algorithms pose a significant challenge in realizing the accurate joint retrieval of these  
22 two parameters at the same location. On this point, a multi-task learning (MTL)  
23 model, which enables the joint retrieval of  $\text{PM}_{2.5}$  concentration and AOD, is proposed  
24 and applied on the top-of-the-atmosphere reflectance (TOAR) data gathered by the  
25 Fengyun-4A Advanced Geosynchronous Radiation Imager (FY-4A AGRI), and  
26 compared to that of two single-task learning (STL) models, namely, random forest  
27 (RF) and deep neural network (DNN). Specifically, The MTL model achieved a  
28 coefficient of determination ( $R^2$ ) of 0.88 and a root mean square error (RMSE) of 0.10  
29 in AOD retrieval. In comparison to the RF model, the  $R^2$  increased by 0.04, the  
30 RMSE decreased by 0.02, and the percentage of retrieval results falling within the  
31 expected error range (Within-EE) rose by 5.55%. The  $R^2$  and RMSE of  $\text{PM}_{2.5}$   
32 retrieval by MTL model are 0.84 and  $13.76\ \mu\text{g}\cdot\text{m}^{-3}$ , respectively. Compared with the  
33 RF model, the  $R^2$  increased by 0.06, the RMSE decreased by  $4.55\ \mu\text{g}\cdot\text{m}^{-3}$ , and the  
34 Within-EE increased by 7.28%. Additionally, compared to the DNN model, the MTL  
35 model showed an increase of 0.01 in  $R^2$  and a decrease of 0.02 in RMSE in AOD  
36 retrieval, with a corresponding increase of 2.89% in Within-EE. For  $\text{PM}_{2.5}$  retrieval,  
37 the MTL model exhibited an increase of 0.05 in  $R^2$ , a decrease of  $1.76\ \mu\text{g}\cdot\text{m}^{-3}$  in

38 RMSE, and an increase of 6.83% in Within-EE. The evaluation suggests that the MTL  
39 model is able to provide simultaneously improved AOD and PM<sub>2.5</sub> retrievals, offering  
40 a critical advantage in capturing the actual distribution of fine particulate matter with  
41 high time efficiency.

42 **Key words:** AOD; PM<sub>2.5</sub>; FY-4A; multi-task learning; joint retrieval

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44 **Article Highlights:**

- 45 ● The simultaneous retrieval of AOD and PM<sub>2.5</sub> concentration at a large spatial scale  
46 is achieved by employing a multi-task learning algorithm.
- 47 ● Multi-task learning can retrieve PM<sub>2.5</sub> and AOD with higher accuracy, compared to  
48 Single-task learning.
- 49 ● Multi-task learning provides new insights on high-efficiency monitoring of aerosol  
50 pollution.

51

## 52 1. Introduction

53 Fine particulate matter ( $PM_{2.5}$ ) refers to suspended particles with aerodynamic  
54 diameter less than or equal to  $2.5\ \mu\text{m}$ , which is an important indicator of air pollution  
55 (Hill et al., 2023). On the other hand, aerosol optical depth (AOD) is a critical  
56 parameter describing the attenuation of solar radiation as caused by aerosols in the  
57 atmosphere. Both  $PM_{2.5}$  and AOD play pivotal roles in air quality, human health (Ho  
58 et al., 2018), and climate change research (Zhang et al., 2017). Therefore, realizing  
59 high-precision detection of aerosols over large areas is important for the environment  
60 and human health. Whereas ground-based measurement constitutes a means of  
61 acquiring reliable and high-precision information, the measurement sites are, more  
62 often than not, distributed unevenly and sparsely. For instance, approximately 90% of  
63  $PM_{2.5}$  monitoring sites in China are concentrated in the eastern and coastal regions,  
64 but in the western regions, such as the Tibetan Plateau or other high-altitude areas,  
65 measurement sites are grossly sparse (Xiao et al., 2016). No better than that, the  
66 existing number of Aerosol Robotic Network (AERONET) sites for AOD  
67 measurements within the Asian-Pacific region is fewer than 90. Therefore,  
68 AERONET AOD data are usually used only for regional studies of aerosols and  
69 validation of remotely sensed data, and effective monitoring of aerosols on a large  
70 scale cannot be realized (Zhang et al., 2018; Hu et al., 2014). The unevenness and  
71 scarcity of ground-based measurement sites, which implies limited observation

72 records, pose significant challenges to conducting comprehensive aerosol studies at a  
73 sub-continental scale, such as over China, which is the country concerning this work.

74 The estimation of surface  $PM_{2.5}$  using satellite AOD data (hereafter, AOD-PM  
75 method) has emerged as the predominant approach over the past decade (Xue et al.,  
76 2020). The mapping from predictors (satellite AOD and auxiliary variables) to the  
77 predictand ( $PM_{2.5}$ ) are mostly data-driven, as exemplified by multiple linear  
78 regression (Gupta & Christopher, 2009), linear mixed-effect model (Lee et al., 2011),  
79 geographically weighted regression (Hu et al., 2013), random forest (RF) (Chen et al.,  
80 2018), convolutional neural network (CNN), long short-term memory (LSTM)  
81 network (Xu et al., 2021; Pak et al., 2018), among other models. An important  
82 premise and theoretical foundation for retrieving surface  $PM_{2.5}$  concentration with  
83 satellite AOD is the strong correlation and connection between  $PM_{2.5}$  and AOD (Li et  
84 al., 2015). Therefore, the availability and reliability of such satellite-derived  $PM_{2.5}$   
85 necessarily depend upon those of satellite AOD. In this regard, to alleviate the errors  
86 introduced by the intermediate AOD estimation, researchers have explored the  
87 possibility of directly obtaining  $PM_{2.5}$  from the top-of-the-atmosphere reflectance  
88 (TOAR) captured by imagers onboard satellites; this approach is known as the  
89 end-to-end method (Yang et al., 2020). For example, Shen et al. (2018) introduced a  
90 TOAR-PM method that established the relationship between TOAR, observation  
91 angles, meteorological factors and  $PM_{2.5}$ ; taking the Wuhan urban agglomeration as  
92 the study area, they achieved an in-sample cross-validated coefficient of

93 determination ( $R^2$ ) and a root mean square error (RMSE) of 0.87 and  $9.89 \mu\text{g}\cdot\text{m}^{-3}$ ,  
94 respectively. Based on an RF model and TOAR data from Himawari-8, Bai et al.  
95 (2021) estimated the  $\text{PM}_{2.5}$  over the Yangtze River delta region; their finding revealed  
96 an  $R^2$  value of 0.75 and a RMSE of  $18.71 \mu\text{g}\cdot\text{m}^{-3}$ . Similarly, Mao et al. (2021)  
97 proposed a RF-based method to directly estimate hourly ground-level  $\text{PM}_{2.5}$  in China  
98 from the Fengyun-4A Advanced Geosynchronous Radiation Imager (FY-4A AGRI)  
99 TOAR, and evaluated all training samples by cross-validation method and keep an  
100 acceptable accuracy ( $R^2 = 0.90$ ,  $\text{RMSE} = 15 \mu\text{g}\cdot\text{m}^{-3}$ ). Yin et al. (2021) also utilized  
101 the Himawari-8 TOAR to estimate the concentration of  $\text{PM}_{2.5}$  over China, but by  
102 employing a light gradient boosting machine (LightGBM); the  $R^2$  and RMSE of  $\text{PM}_{2.5}$   
103 estimated are 0.83 and  $23.7 \mu\text{g}\cdot\text{m}^{-3}$ .

104 As for AOD, several physical retrieval algorithms have long been well known,  
105 including the dark target (DT) (Kaufman et al., 1997), deep blue (DB) and  
106 Multi-Angle Atmospheric Correction (MAIAC) algorithms (Lyapustin et al., 2018).  
107 Similar to the case of  $\text{PM}_{2.5}$ , end-to-end AOD estimation algorithms have also been  
108 used as promising alternatives to the physical algorithms. For instance, Ding et al.  
109 (2022) employed a so-called “neural network aerosol retrieval for geostationary  
110 satellite (NNAeroG)” method to estimate AOD over the full disk area of FY-4A. The  
111 validation result demonstrated an RMSE of 0.24, an  $R^2$  of 0.73, and 58.7% AOD  
112 values falling within an expected error envelope of  $\pm (0.05 + 15\% \times \text{AOD}_{\text{AERONET}})$ ,  
113 which is abbreviated as EE15 hereafter. She et al. (2022) proposed a Landsat-8 AOD

114 retrieval algorithm based on DNN, which successfully retrieved AOD in a longitude  
115 range of 30°W–160°E and a latitude range of 60°S–60°N. The estimated AOD was  
116 found to exhibit excellent agreement with AERONET AOD, with an  $R^2$  of 0.71, an  
117 RMSE of 0.15, and 61% of the retrieved AOD values falling within an expected error  
118 envelope of  $\pm (0.05 + 20\% \times \text{AOD}_{\text{AERONET}})$ .

119       Regardless of the retrieval subject (i.e.,  $\text{PM}_{2.5}$  or AOD), most current  
120 TOAR-based retrieval methods are single-task learning (STL) models. In STL,  
121 individual tasks employ independently trained models without explicit mechanisms to  
122 support information sharing across different tasks. However, applying STLs on  
123 TOAR data, though simple in conception, are confined by many factors. For instance,  
124 there is a conceptual difference between  $\text{PM}_{2.5}$ , which reflects the near-surface  
125 turbidity of the atmosphere, and TOAR, which covers the atmospheric information  
126 from surface to several hundred kilometers in altitude. This vertical distribution depth  
127 mismatch makes  $\text{PM}_{2.5}$  retrieval more challenging. On the other hand, retrieving AOD  
128 from TOAR faces two general difficulties. First, ground-based observatories based on  
129 ground-based instruments (e.g., sun photometers) are mainly located in the eastern  
130 and coastal regions of China, such as the North China Plain region, the Yangtze River  
131 Delta region, and the Pearl River Delta region (Xun et al., 2021). Due to the sparse  
132 distribution of these stations, data-driven models may not be able to capture more  
133 complex regional features, thus risking overfitting. Second, missing values often  
134 occur due to cloud cover (Kokhanovsky et al., 2007), thereby reducing the number of

135 available samples. During the cloud detection process, heavy aerosols can be  
136 misclassified as clouds (Song et al., 2019), which enhances the mutual interference  
137 between clouds and aerosols. Therefore, there are significant limitations in improving  
138 the retrieval accuracy of  $PM_{2.5}$  and AOD. It is encouraging that multiple studies have  
139 shown a correlation between  $PM_{2.5}$  and AOD (Zheng et al., 2017; Yang et al., 2019).  
140 Given the existence of this correlation, is it possible to employ a method that  
141 effectively utilizes this relationship to jointly retrieve  $PM_{2.5}$  and AOD using TOAR  
142 data, even with a limited number of samples?

143 In what follows, any method that allows for the joint retrieval of multiple  
144 correlated parameters is referred to as multi-task learning (MTL). One distinct  
145 advantage of MTL over STL is that it allows parameters sharing to a certain degree  
146 between several related learning tasks, thereby improving the performance of all tasks  
147 (Ruder, 2017; Zhang & Yang, 2017). In fact, MTL is a general concept that goes far  
148 beyond the retrieval of atmospheric parameters. But in recent years, several MTL  
149 approaches have been adopted by atmospheric scientists. For instance, Zhang et al.  
150 (2020) proposed an MTL model that combines CNN and gated recurrent unit (GRU)  
151 for multi-step-ahead multi-station  $PM_{2.5}$  prediction. It is tested on three monitoring  
152 stations located in three different districts of Lanzhou, China, with an RMSE of 7.96  
153  $\mu\text{g}\cdot\text{m}^{-3}$ , indicating better performance in intensive air quality prediction than previous  
154 models based on simple hybridization. Further, to fully utilize the meteorological  
155 information from the monitoring stations, Xu and Yoneda (2021) proposed a long



156 short-term memory (LSTM) autoencoder MTL model to predict PM<sub>2.5</sub> time series in  
157 multiple locations city wide, which greatly improved the prediction accuracy and  
158 calculation cost compared with the traditional LSTM model. Meanwhile, in order to  
159 verify the performance difference between MTL model and STL model, Song et al.  
160 (2022) employed attentive MTL model to predict air quality in urban stations. In the  
161 comparison of results in Seoul, the proposed attentive model with MTL outperformed  
162 the STL attentive model in terms of accuracy performance, with RMSE values for  
163 PM<sub>2.5</sub> being 8.36 and 8.66  $\mu\text{g}\cdot\text{m}^{-3}$ , respectively. On the other hand, in order to  
164 enhance the input, the spectral and spatial information is jointly used to retrieve fine  
165 mode fraction (FMF), Chen et al., (2020) proposed an artificial neural network for  
166 aerosol retrieval (NNAero) to jointly retrieve AOD and (FMF). The input data are the  
167 Moderate Resolution Imaging Spectroradiometer (MODIS) TOAR together with  
168 MODIS-derived surface reflectance in 5 spectral bands, and the labels are AERONET  
169 FMF and AOD. The results show that 68% of the NNAero AOD values are within  
170 EE15, which is better than the DT algorithm (31% within EE15).

171 This contribution aims to simultaneously retrieve AOD and PM<sub>2.5</sub> concentrations  
172 over China using FY-4A AGRI data based on an MTL algorithm. The performance of  
173 the model is thoroughly evaluated using ground-based observations and compared  
174 with that of the classical RF and DNN models. The rest of the paper is organized as  
175 follows: Section 2 introduces the various datasets used in this study, among which the  
176 FY-4A AGRI data and European Center for Medium-Range Weather Forecasts'

177 (ECMWF's) fifth-generation reanalysis (ERA5) are used as model inputs, whereas the  
178 Multi-Angle Atmospheric Correction (MAIAC) AOD data and the ground-based  
179 PM<sub>2.5</sub> observations from two sources act as the training targets and out-of-sample  
180 verifications. In Section 3, a detailed description of data processing and the setups of  
181 the machine-learning models are provided. Section 4 summarizes the statistical results  
182 and evaluates the performance of the models. Finally, conclusion follows in Section 5.

## 183 **2. Data**

184 This study utilized FY-4A AGRI and ECMWF ERA5 data, spanning from  
185 March 12, 2018, to March 11, 2019, covering the geographical expanse of China, as  
186 inputs for the MTL model. For the same region and time frame, the MAIAC AOD  
187 and the PM<sub>2.5</sub> measurements, collected by the China Environmental Monitoring  
188 Center (CEMC), are used as the targets during learning. Furthermore, PM<sub>2.5</sub>  
189 measurements obtained from the Environmental Protection Department (EPD) in  
190 Hong Kong (HK) and AERONET AOD data are utilized for out-of-sample  
191 verification. The Long-term Gap-free High-resolution Air Pollutant (LGHAP)  
192 concentration dataset was employed to assess and validate the performance of the  
193 MTL model in Northwest China.

### 194 ***2.1 FY-4A /AGRI data***

195 FY-4A, which was launched on December 11, 2016, has capabilities that are  
196 greatly enhanced, as compared to its predecessors, in terms of environmental and

197 weather monitoring, warning, and forecasting. The AGRI sensor onboard FY-4A has  
198 14 channels and a wavelength range of 0.45–13.8  $\mu\text{m}$ , covering the visible (VIS),  
199 near-infrared (NIR), medium infrared, and long infrared spectra, with a spatial  
200 resolution of 1km (Fu et al., 2024). The level-1 FY-4A AGRI TOAR data are  
201 available online (<http://satellite.nsmc.org.cn/>). In this study, three channels that are  
202 known to related to aerosol are selected, namely, 0.45–0.49  $\mu\text{m}$  (CH01), 0.75–  
203 0.90  $\mu\text{m}$  (CH03), and 2.1–2.35  $\mu\text{m}$  (CH06). Furthermore, geometrical properties  
204 including solar zenith angle (SOZ), solar azimuth angle (SOA), satellite zenith angle  
205 (SAZ), and satellite azimuth angle (SAA) are also computable parameters that are to  
206 be used as inputs.

## 207 ***2.2 Meteorological data***

208 ERA5, which is a global atmospheric reanalysis product developed by ECMWF  
209 with a spatial resolution of  $0.25^\circ \times 0.25^\circ$ , is used as an auxiliary dataset to provide the  
210 meteorological factors that can affect the composition and transport of aerosols (Deng  
211 et al., 2012). The selected meteorological factors include the u- and v-component  
212 wind at a height of 10 m (U10, V10, m/s), temperature at a height of 2 m (T2m, K),  
213 surface pressure (Sp, Pa), boundary layer height (Blh, m), total column ozone (Tco3,  
214 DU), total column water vapor (Tcww,  $\text{kg}/\text{m}^2$ ); these can be directly downloaded from  
215 the C3S climate data store (<https://cds.climate.copernicus.eu/>) with relative ease,  
216 owing to the mature data dissemination system of ECMWF.

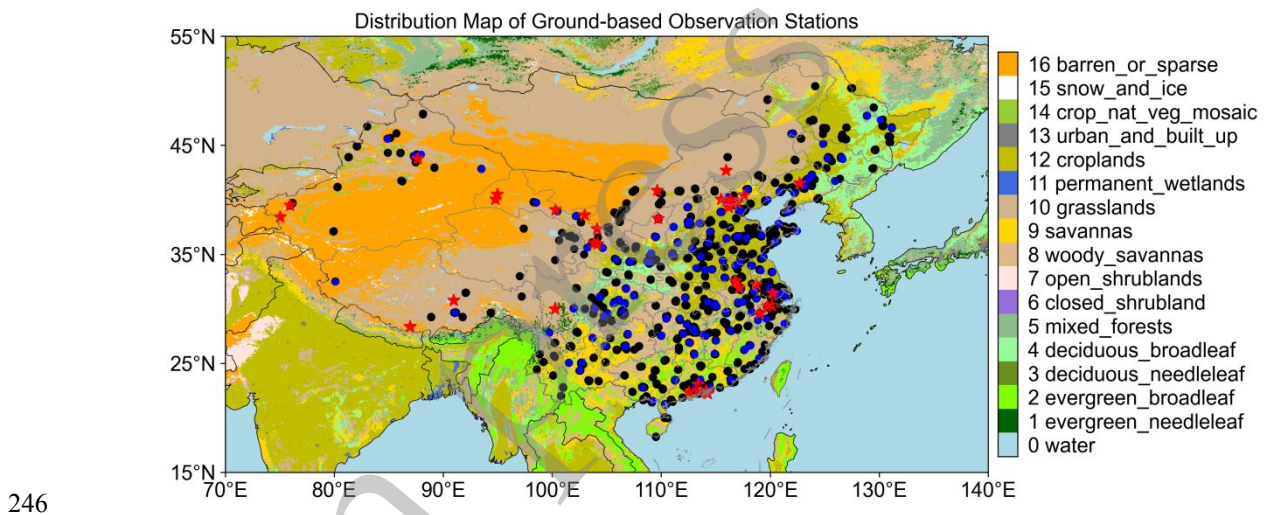
### 217 **2.3 MODIS MAIAC AOD**

218 The Terra and Aqua satellites, with MODIS sensors onboard, were launched in  
219 December 1999 and May 2002, respectively. There are four standard land MODIS  
220 AOD products, which have been retrieved using DT, DB, DT-DB combined, and  
221 MAIAC algorithms (Xie et al., 2019). Among these products, the MAIAC AOD has  
222 the highest spatial resolution of 1 km. Validation results with AERONET AOD show  
223 that MAIAC AOD retrievals are highly correlated with ground-based AOD  
224 measurements. The correlation coefficients (R) are greater than 0.8 at more than 68%  
225 of AERONET sites. The accuracy of MAIAC AOD retrievals is high (within expected  
226 error (EE) = 87.49% and 83.15%) in the regions of tropical rainforest climate and  
227 tropical open forest climate (Qin et al., 2021). For that reason, the daily MAIAC AOD  
228 product at the wavelength of 550 nm is used in this work. This product has passed  
229 quality assurance, cloud screening, and adjacency testing (Lyapustin et al., 2018).

### 230 **2.4 Ground-based observation data**

231 Quality-controlled hourly PM<sub>2.5</sub> concentration measurements are collected from  
232 CEMC (<http://www.cnemc.cn>), with regularly calibrated sensors located at 1590 sites  
233 across the China (see Fig. 1); these data are to be used as training targets. For  
234 out-of-sample verification, the hourly PM<sub>2.5</sub> concentrations data at 15 air-quality  
235 monitoring stations in HK are acquired from the Environmental Protection  
236 Department (EPD) website (<http://www.epd.gov.hk/epd/>). The hourly PM<sub>2.5</sub>  
237 concentration were measured using the micro-oscillation balance method and  $\beta$

238 absorption method, and the uncertainty is said to be within  $5 \mu\text{g}\cdot\text{m}^{-3}$  (Miao and Liu,  
239 2019). As for AOD data, they are sources from the AERONET, which is the world's  
240 largest ground-based aerosol observation network. AERONET provides observations  
241 of aerosol properties (<https://aeronet.gsfc.nasa.gov/>), such as optical, microphysical or  
242 radiative properties. AERONET provides AOD in seven wavelength bands (340, 380,  
243 440, 500, 675, 870, and 1020 nm). In this study, AOD at 550 nm is least-squares fitted  
244 using the well-known quadratic relationship between AOD and wavelength (Fu et al.,  
245 2023).



247 **Fig. 1.** Spatial distribution of AERONET sites (red stars) and  $\text{PM}_{2.5}$  ground stations  
248 (blue/black dots). Annotated with black and blue dots are  $\text{PM}_{2.5}$  sites used for model  
249 training and testing, respectively. The background map indicates the type of ground  
250 cover from MODIS in 2018.

## 251 **2.5 LGHAP concentration dataset**

252 We obtained the LGHAP dataset from the Earth System Science Data website

253 (<https://www.earth-system-science-data.net>). The dataset synergistically integrates  
254 multimodal aerosol data from different sources using a tensor-flow-based data fusion  
255 method to generate daily gapless AOD products for China from 2000 to 2020 with a  
256 spatial resolution of 1 km. Subsequently, the PM<sub>2.5</sub> concentration on continuous space  
257 was estimated using an integrated learning approach. The data were stored in NetCDF  
258 format, and code was provided to help users read and visualize the data (Bai et al.,  
259 2022). The verification results of ground observation data show that LGHAP AOD  
260 data is highly correlated with AERONET AOD data, with an R of 0.91 and RMSE  
261 equaling 0.21. In addition, PM<sub>2.5</sub> estimates were highly correlated with ground-based  
262 measurements, with an R of 0.95 and RMSE equaling 12.03  $\mu\text{g}\cdot\text{m}^{-3}$ . In this study, we  
263 used LGHAP AOD and LGHAP PM<sub>2.5</sub> data to verify the estimation results of MTL  
264 model in Northwest China.

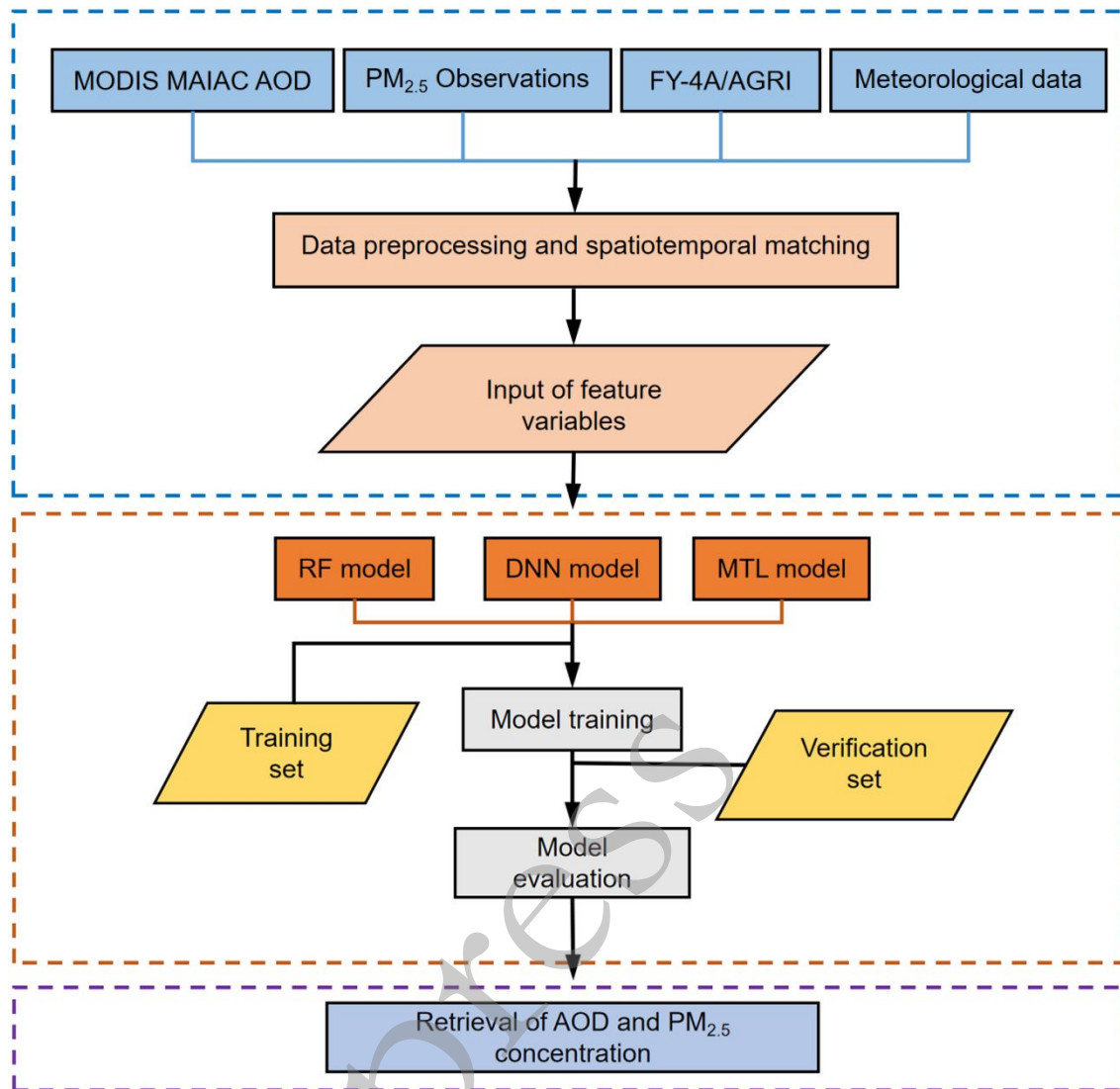
### 265 **3 Methodology**

266 Figure 2 illustrates the workflow of this study, including dataset preprocessing,  
267 model training and validation, and retrieval of AOD and PM<sub>2.5</sub> concentrations. The  
268 workflow is applied not only to the MTL model but also two classical STL models  
269 (RF and DNN).

#### 270 **3.1 Data preprocessing**

271 Because data from multiple sources are used in the study, in order to ensure  
272 spatial scale consistency, satellite and auxiliary reanalysis data are resampled to a grid

273 size of  $0.01^\circ \times 0.01^\circ$  using bilinear interpolation (Wei et al., 2019). Subsequently, all  
274 data are spatially matched based on the nearest-neighbor grid cells to the  $PM_{2.5}$   
275 stations. Regarding temporal matching,  $PM_{2.5}$  values from two hours adjacent to  
276 MODIS overpass time (02:30 UTC for Terra and 05:30 UTC for Aqua) and FY-4A  
277 data within 2.5 min of the overpass time are averaged. Reanalysis data meeting this  
278 temporal requirement are drawn from the time series. To filter for data with high  
279 quality and reliability, samples containing zero and invalid values are rejected. At the  
280 same time, because  $PM_{2.5}$  and AOD have lognormal distributions, a logarithmic  
281 transformation is applied, as to make the data more symmetric and stable. After data  
282 filtering, a total of 132,540 samples remains for model training and validation. Of  
283 these, 90% of the sites are used for training the model (marked as black dot in Fig.1),  
284 while the remaining are used for testing the model (marked as red dot in Fig. 1). It is  
285 worth noting that the proportion of MODIS MAIAC AOD data divisions based on  
286  $PM_{2.5}$  sites matched is the same as for  $PM_{2.5}$ . To mitigate the impacts of outliers on  
287 model fitting, normalization and standardization must be applied to all variables prior  
288 to training. In the feature selection stage, we adopted the method of permutation  
289 feature importance to select 17 most representative features as the input of the model,  
290 as shown in Table 1. Permutation feature importance is a model inspection technique  
291 that measures the contribution of each feature to a fitted model's statistical  
292 performance on a given tabular dataset (<https://scikit-learn.org/>).



293

294 **Fig. 2.** Machine learning framework to retrieve AOD and PM<sub>2.5</sub> concentrations from  
 295 various sites in China.

296



297 **Table 1.** Details of input datasets used in this study.

Field	Variables	Spatial resolution (km)	Main scientific objectives
FY-4A AGRI	0.45–0.49um (CH01)	1	Small particle aerosol, true color
	0.75–0.90um (CH03)	1	Vegetation, aerosols
	2.1–2.35um (CH06)	2-4	Cirrus cloud, aerosol, particle size
FY-4A geo-data	SAA	4	Position information from ground stations and satellite
	SAZ	4	
	SOA	4	
	SOZ	4	
	Lon	-	
	Lat	-	
ECMWF/ERA5	Blh	25	Planetary boundary layer height
	T2m	25	2-m air temperature
	U10	25	10-m eastward wind
	V10	25	10-m northward wind
	Tcwv	25	Total column water vapor
	Tco3	25	Total column ozone
	Sp	25	Surface pressure
Others	Mon	-	The month values were converted to a cosine distribution in the range of -1 to 1 to better represent seasonal effects.

298

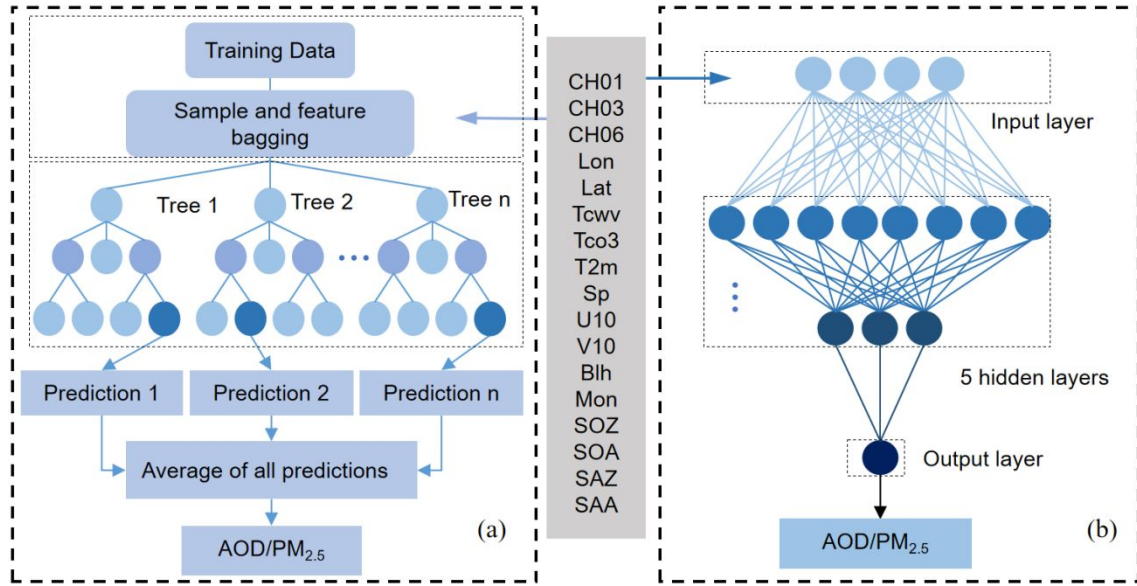
299 **3.2 STL models**

300 In the field of satellite remote sensing, RF and DNN are two widely used  
301 machine-learning algorithms. RF is a popular ensemble learning method that has been  
302 widely applied for the retrieval of surface solar radiance (Shi et al., 2023), PM<sub>2.5</sub> and  
303 AOD estimation (Wei, et al., 2019; Chen et al., 2018). Similarly, She et al. (2020)

304 demonstrated that AOD retrieved with a DNN model has good reliability. Given such  
305 previous experience, RF and DNN are selected in this study as the STL models for  
306 benchmarking purposes.

307 RF is a classic ensemble learning method based on decision trees, which can  
308 quantify nonlinear relationships; its network structure is depicted in Fig. 3a. The  
309 algorithm has good robustness and the ability to deal with high-dimensional data and  
310 can deal with missing values and outliers effectively. In addition, the output results  
311 are easily interpretable and understandable.

312 The DNN is simply a multilayer perceptron with more than two hidden layers  
313 (Yuan et al., 2020). The DNN architecture used in this paper is presented in Fig. 3b,  
314 which consists of one input layer (17 predictors), five hidden layers, and one output  
315 layer, which outputs either AOD or  $PM_{2.5}$  depending on the learning task. Each  
316 hidden layer contains neurons, whose value is computed by taking a linear  
317 combination of all the neuron values from the previous layer using multiple weights  
318 and a bias term; in that the network is a fully connected one. This is followed by the  
319 application of a nonlinear activation function, to make the output more relevant to the  
320 predicted value. In this study, the rectified linear unit (ReLU) is chosen as the  
321 activation function (Glorot et al., 2011).



322

323 **Fig. 3.** The architecture of (a) random forest algorithm and (b) deep neural network.

### 324 3.3 MTL model

325 To incorporate the potential nonlinear spatial correlations between AOD and  
 326 PM<sub>2.5</sub> retrievals, MTL is thought functional, as it is able to learn multiple related  
 327 prediction tasks at the same time and to share the feature information of multiple tasks  
 328 (Ruder, 2017). Compared with STL, MTL has stronger abilities in generalization and  
 329 feature learning, and thus has become increasingly popular in the field of artificial  
 330 intelligence (Ranjan et al., 2019). When performing MTL, there lie two main  
 331 difficulties: handling multiple loss functions and designing an efficient parameter  
 332 sharing mechanism.

333 When dealing with the loss function during MTL, a common practice is to add  
 334 each individual loss function linearly, as illustrated in Eq. (1):

$$335 L_{\text{total}} = \sum w_i L_i \quad (1)$$

336 where  $w_i$  and  $L_i$  represent the weights and losses of task  $i$ , respectively, and  $L_{\text{total}}$

337 represents the total loss function. Since the overall performance heavily relies on the  
338 weights assigned to each loss function, traditional weighing methods lack a reliable  
339 basis and adjusting these weights can be time-consuming and laborious, which  
340 subsequently making it difficult to achieve the desired optimization result. In addition,  
341 there often exist conflicting objectives among multiple tasks, such as improving the  
342 performance of one task may lead to a decline in the performance of another task,  
343 which makes finding an effective set of weights to achieve simultaneous optimization  
344 of two tasks accompanied by high uncertainty.

345 Speaking of uncertainty, MTL is usually associated with both the cognitive  
346 uncertainty and accidental uncertainty. More specifically, cognitive uncertainty is  
347 caused by the model itself, where the outcome to be predicted is beyond the knowing  
348 range of the model; however, this uncertainty may be reduced as the training data  
349 points get more numerous. Accidental uncertainty includes heteroscedastic and  
350 homoscedastic uncertainties. Whereas the former refers to the uncertainty caused by  
351 differences in the input data, also known as data dependence, the latter is commonly  
352 used to characterize the data noise between different tasks, in that, the optimal weight  
353 for each task depends on its noise size (Kendall et al., 2018). In view of that, the main  
354 consideration, as to the weight assignment of the two learning tasks, is the  
355 homoscedastic uncertainty. In short, the total loss function is assumed to take the  
356 form:

357 
$$L_{(w, \sigma_1, \sigma_2)} = \frac{1}{2\sigma_1^2}L_1(w) + \frac{1}{2\sigma_2^2}L_2(w) + \log \sigma_1 \sigma_2 \quad (2)$$

358 where  $\sigma_1$  and  $\sigma_2$  signify uncertainties arising from data noise, serving as learnable  
359 parameters. As  $\sigma$  increases, the associated weight decreases; conversely, with  
360 decreasing noise  $\sigma$ , the corresponding weight increases. To mitigate overfitting,  
361 regularization terms are appended to the end of the loss function. During training,  
362  $\log \sigma_1 \sigma_2$  is introduced as a trainable variable, effectively constraining the loss  
363 function's variation range and preventing division by zero anomalies. In this study, the  
364 initial value of  $s$  is set as a random number greater than 0. As the number of training  
365 increases, the  $s$  representing the noise characteristics of  $PM_{2.5}$  and AOD eventually  
366 converges to 3.7 and 0.7.

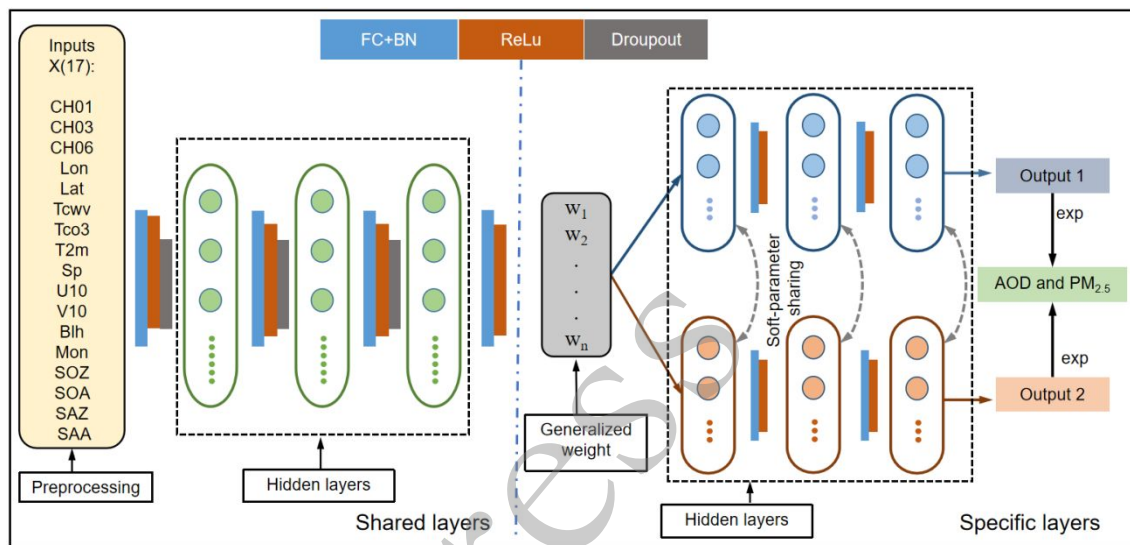
367 Parameter sharing in MTL has two main mechanisms: hard parameter sharing  
368 and soft-parameter sharing (Sun et al., 2019). Hard parameter sharing is suitable for  
369 tasks with strong correlation and can effectively reduce the risk of overfitting.  
370 However, when the correlation of outputs between tasks is poor, MTL with only hard  
371 parameter sharing may not fully satisfy the requirements of all tasks and deteriorate  
372 model performance instead. On the other hand, soft-parameter sharing allows each  
373 task to have separate parameters and hidden layers, while still enables information  
374 access among tasks. The way to achieve this is to regularize the loss at the output  
375 layer, by measuring the distance between the models (Maurer et al., 2012). The  
376 distance, as typically gauged with  $L_1$  or  $L_2$  loss, describes the parameter similarity  
377 between the same layer of different models, thereby encouraging the similarity of  
378 parameters across multiple tasks. This strategy achieves the goal of preventing

379 complete independence in the specific layer and increasing the utilization of pertinent  
380 information.

381 The structure of MTL used in the study is shown in Fig. 4, which contains four  
382 shared layers and three specific layers for each task. Each layer is fully connected to  
383 the adjacent ones, with other techniques and settings including batch normalization,  
384 ReLU activation and dropout mechanism implemented for the network. Batch  
385 normalization seeks to accelerate convergence during training by normalizing the  
386 inputs in each small training batch of data. The shared layers adopt hard parameter  
387 sharing, which ensures that each neuron has identical parameters for both tasks.  
388 Consequently, a set of feature weights ( $w_1, w_2, \dots, w_n$ ) that exhibit generalization  
389 performance is obtained at the end of shared layers. The next steps involve specific  
390 layers. In these specific layers, each task is equipped with an independent model,  
391 possessing its own set of parameters. During training, the parameters of each model  
392 corresponding to each task are updated independently. Following each update, the  
393 distance between model parameters (L2 norm) is introduced as a regularization term  
394 to ensure similarity among parameters as much as possible. This notion of  
395 soft-parameter sharing is greatly inspired by MTL regularization techniques.

396 In addition to setting up the loss function and establishing the model structure,  
397 the selection of hyperparameters is also very important. For example, if the learning  
398 rate is set too high, then the parameter update may be too drastic, causing the loss  
399 function to not converge effectively; Larger batch sizes improve the stability of

400 parameter updates. In the training process, we used a 10-fold cross-validation method  
 401 to select the best hyperparameters of the MTL model. Specifically, we set the number  
 402 of iterations, learning rate and batch size to 378, 0.001 and 1000, respectively, and  
 403 selected the Adam optimizer to optimize the weight parameters in the neural network  
 404 model.



405

406 **Fig. 4.** The architecture of MTL model. Both shared layers (left) and specific layers  
 407 (right) are full-connected (FC). Hard parameter sharing is used in shared layers while  
 408 soft-parameter sharing is applied to the specific layers. ReLU is the activation layer,  
 409 BN is a batch normalization layer.

### 410 3.4 Model Validation

411 A 10-fold cross-validation method is applied in the training dataset (recall  
 412 Section 3.1) to select the best model parameters (Rodríguez et al., 2010). The testing  
 413 dataset is used to evaluate the model performance based on statistical metrics  $R^2$ ,  
 414 RMSE, and mean absolute error (MAE). The EE is used to evaluate the accuracy of

415 the MTL model, and the calculation formula is:

$$416 \quad EE = (1 \pm 0.15)y_i \pm b_i \quad (3)$$

417 where  $y_i$  represents the label value of ground-based AOD or  $PM_{2.5}$ , and  $b_i$   
418 represents the intercept, which takes the value of 0.05 for AOD verification and 5  
419  $\mu\text{g}\cdot\text{m}^{-3}$  for  $PM_{2.5}$  verification.

## 420 **4 Results and discussion**

### 421 **4.1 Comparison among three models**

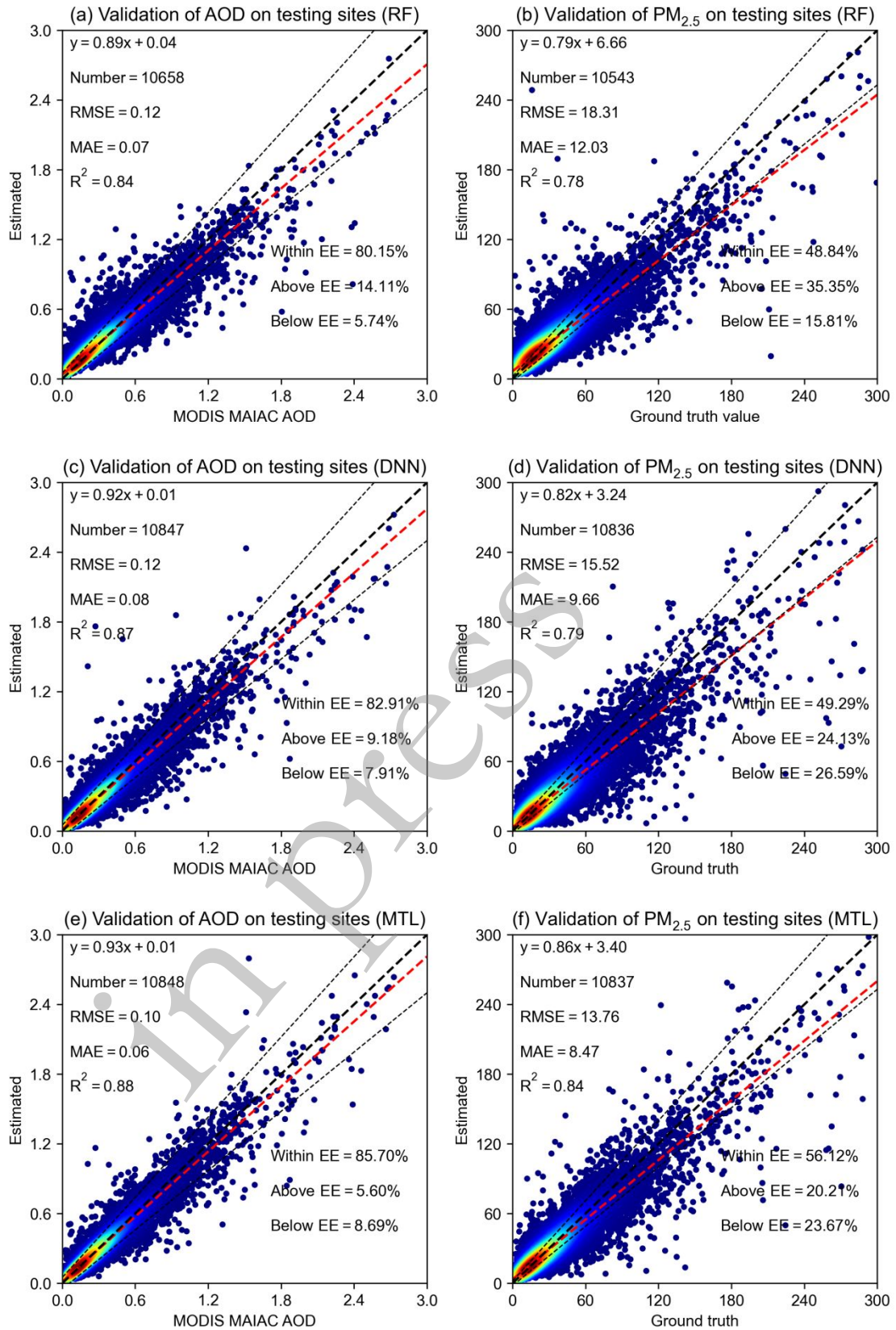
422 Here, both STL models and MTL models are validated with test data. First,  
423 according to the spatial distribution of ground stations, 10% of  $PM_{2.5}$  observation  
424 stations (159 blue dots in Fig. 2) are uniformly selected as the test data set  
425 (out-of-station validation), and through spatio-temporal matching between these  
426 stations and MODIS MAIAC AOD, the AOD verification data set at the test station is  
427 obtained. The parameters used to evaluate model performance are  $R^2$ , RMSE, MAE,  
428 and EE.

429 Figure 5a displays the AOD predictions by the RF model, with respect to  
430 MAIAC AOD. The model has an  $R^2$  of 0.84, an RMSE of 0.12, and an MAE of 0.07,  
431 with 80.15% of Within-EE predictions. However, the RF model tends to overestimate  
432 AOD for low MAIAC AOD values and underestimate it for high MAIAC AOD  
433 values. The DNN model (Fig. 5c) sees improvements over the RF model, with an  $R^2$   
434 of 0.87. However, there are no significant changes in RMSE and MAE (0.12 and



435 0.08). The MTL model (Fig. 5e) performs the best, with the evaluation metrics taking  
436  $R^2 = 0.88$ ,  $RMSE = 0.10$ ,  $MAE = 0.06$ , and 85.70% Within-EE predictions.

437 In terms of the  $PM_{2.5}$  prediction, results of STL model and MTL model are  
438 compared. It can be seen from Fig. 5b and Fig. 5d that the result of DNN-STL ( $R^2 =$   
439  $0.79$ ,  $RMSE = 15.52 \mu\text{g}\cdot\text{m}^{-3}$ ,  $MAE = 9.66 \mu\text{g}\cdot\text{m}^{-3}$ ) is better than that of RF-STL ( $R^2 =$   
440  $0.78$ ,  $RMSE = 18.31 \mu\text{g}\cdot\text{m}^{-3}$ ,  $MAE = 12.03 \mu\text{g}\cdot\text{m}^{-3}$ ), but the optimization effect is not  
441 obvious, Within-EE only increased by 0.45%. The  $R^2$ ,  $RMSE$ ,  $MAE$ , and Within-EE  
442 for the MTL model are 0.84,  $13.76 \mu\text{g}\cdot\text{m}^{-3}$ ,  $8.47 \mu\text{g}\cdot\text{m}^{-3}$  and 56.12% (Fig. 5f),  
443 respectively, with significant improvement in all metrics compared with the STL  
444 model. Compared with the traditional STL model, the MTL model developed in this  
445 study optimized both tasks (AOD and  $PM_{2.5}$ ) to a certain extent. The results further  
446 show that the MTL model is more effective in optimizing  $PM_{2.5}$  than AOD.



447

448 **Fig. 5.** Validation of estimated AOD and  $PM_{2.5}$  on testing sites of RF (a, b), DNN (c,

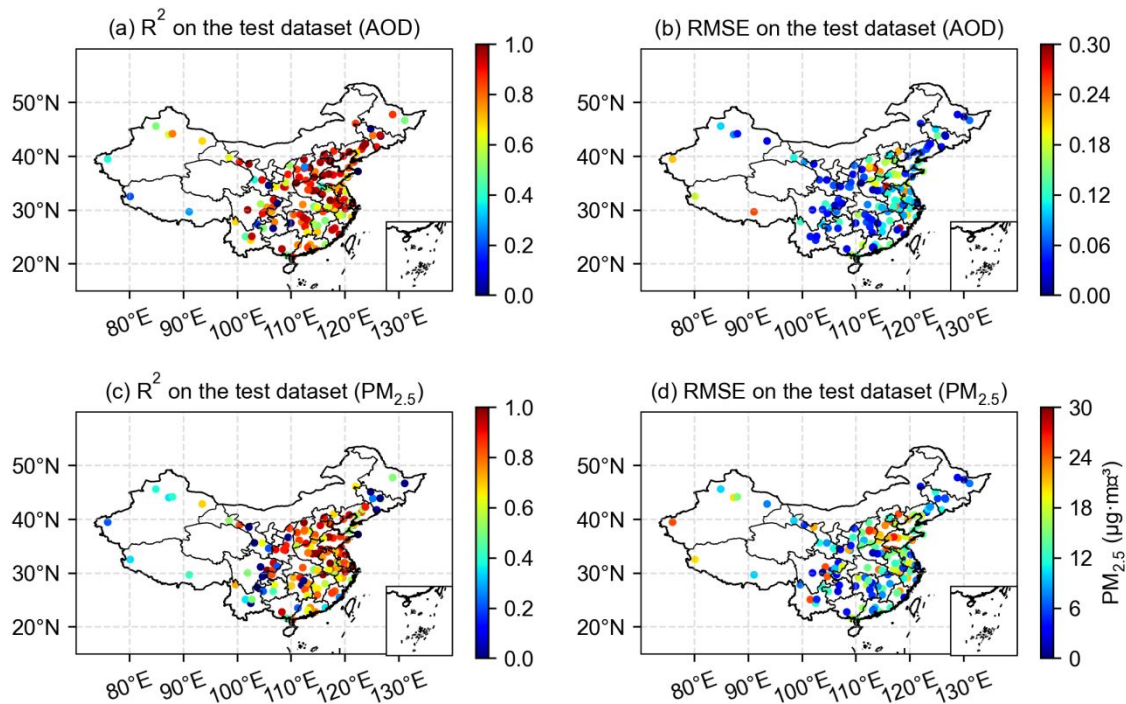
449 d), and MTL model (e, f). The light dotted line represents the EE line, the dark dashed  
450 line represents the 1:1 line, and the red dashed line is the linear regression fitting line.

#### 451 ***4.2 Site-specific model performance***

452 Figure 6 illustrates the spatial distributions of the four-evaluation metrics for the  
453 MTL model on testing sites, revealing significant regional disparities and trends. In  
454 most sites across eastern and coastal China, the values of  $R^2$  exceed 0.8 (Fig. 6a and  
455 Fig. 6c), indicating its capability to effectively capture variations in AOD and  $PM_{2.5}$   
456 with a high accuracy. However, the performance of the MTL model experiences a  
457 substantial decline in western China ( $R^2$  values are generally less than 0.5),  
458 particularly in areas characterized by complex topography like the Tibetan Plateau.  
459 This observation can be primarily attributed to the scarcity of ground-based  
460 observation sites in that region, hindering the ability of the model to adapt to the  
461 unique aerosol patterns prevalent in this distinctive environment (Fang et al., 2016).

462 Moreover, AOD and  $PM_{2.5}$  exhibit a pronounced north-to-south gradient in terms  
463 of their RMSE distribution in China (Fig. 6b and Fig. 6d), which is consistent with the  
464 findings of previous studies (Wei, et al., 2019). Specifically, sites in the North China  
465 Plain (NCP), which is an area known for its significant anthropogenic emissions, and  
466 Northwest China, which is characterized by substantial natural dust emissions (Gui et  
467 al., 2020), show higher RMSE values for AOD and  $PM_{2.5}$ , exceeding 0.18 and 18  
468  $\mu\text{g}\cdot\text{m}^{-3}$ .

469



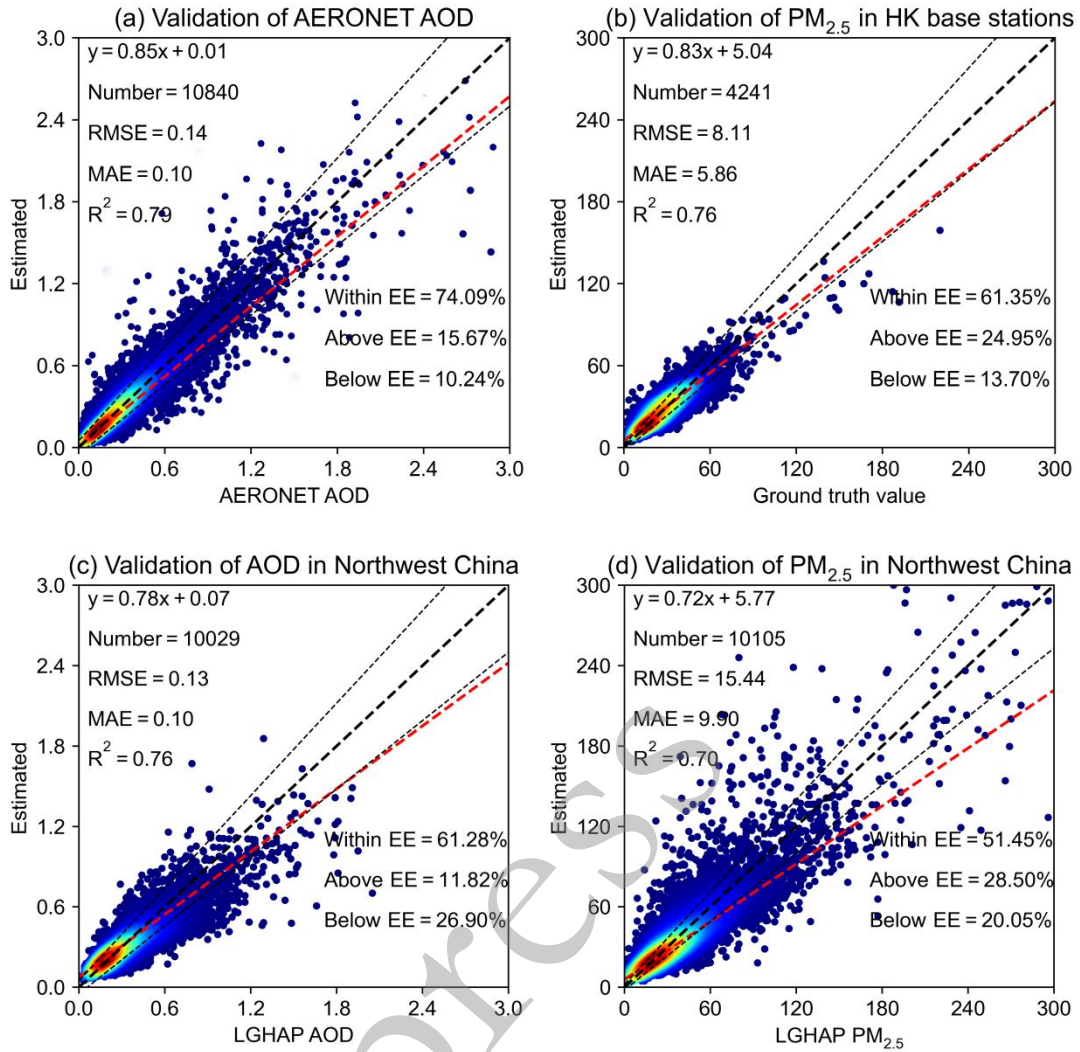
470

471 **Fig. 6.** Spatial distribution of MTL model evaluation metrics (a, c:  $R^2$  for AOD and  
 472  $PM_{2.5}$ , b, d: RMSE for AOD and  $PM_{2.5}$ .)

473 In order to comprehensively assess the discrepancy between the estimation  
 474 results of the MTL model and the ground-based observations, we utilized the AOD  
 475 and  $PM_{2.5}$  ground-based observations for the period of March 12, 2018 to March 11,  
 476 2019 to validate the accuracy of the MTL model. These data include AOD  
 477 observations from 38 AERONET stations in China and  $PM_{2.5}$  data from 15 air quality  
 478 monitoring stations in Hong Kong. Fig 7a demonstrates the comparison results  
 479 between the AERONET AOD and the MTL AOD, with an  $R^2$  of 0.79, an RMSE of  
 480 0.14, an MAE of 0.10, and a Within-EE of 74.09%. In addition, Fig 7b demonstrates  
 481 the results of the comparison between the MTL  $PM_{2.5}$  and the ground-level  $PM_{2.5}$   
 482 observations in Hong Kong, with an  $R^2$  of 0.76, an RMSE of  $8.11 \mu\text{g}\cdot\text{m}^{-3}$ , an MAE of

483  $5.86 \mu\text{g}\cdot\text{m}^{-3}$ , and a Within-EE of 61.35%. These results show that the predictions of  
484 the MTL model are in good agreement with the ground observations, verifying the  
485 reliability of the MTL model.

486 Due to the limited number of ground-based observation sites in Northwest China,  
487 it is difficult to provide sufficient data for validating the performance of the MTL  
488 model in the region. Therefore, in this study, observation sites distributed in 28 cities  
489 in the region were selected and LGHAP AOD and LGHAP  $\text{PM}_{2.5}$  data from March  
490 12, 2018, to March 11, 2019, were obtained by spatio-temporal matching. In Fig. 7c,  
491 the results of the comparison between MTL AOD and LGHAP AOD are  
492 demonstrated, with  $R^2$  of 0.76, RMSE of 0.13, MAE of 0.10, and Within-EE of  
493 61.28%. And in Fig. 7d, the results of comparison between MTL  $\text{PM}_{2.5}$  and LGHAP  
494  $\text{PM}_{2.5}$  are demonstrated with  $R^2$  of 0.70, RMSE of  $15.44 \mu\text{g}\cdot\text{m}^{-3}$ , MAE of  $9.90 \mu\text{g}\cdot\text{m}^{-3}$ ,  
495 and Within-EE of 51.45%. The results show that the validation results of the MTL  
496 model with ground-based observation sites are better than the validation results of the  
497 MTL model with the LGHAP AOD and LGHAP  $\text{PM}_{2.5}$  in Northwest China,  
498 especially in terms of  $\text{PM}_{2.5}$ . This is mainly due to the limited training data in the  
499 region and the fact that most of the region is in plateaus, basins and mountains with  
500 cold and dry winters and high summer temperatures. In particular, the Taklamakan  
501 Desert surrounded by high mountains and the Gobi Desert in Inner Mongolia are one  
502 of the dust source areas in China, which makes the aerosol types in Northwest China  
503 more complex and diverse.



504

505 **Fig. 7.** Scatter plot of validation results of MTL AOD and  $PM_{2.5}$  in different regions.

506 (a) Comparison results of AERONET AOD and MTL AOD. (b) Comparison results

507 of HK Ground-based Observations  $PM_{2.5}$  and MTL  $PM_{2.5}$ . (c), (d): The results of MTL

508 AOD ( $PM_{2.5}$ ) and LGHAP AOD ( $PM_{2.5}$ ) were compared in Northwest China.

### 509 4.3 Feature importance of MTL model

510 Owing to the “black box” nature of ML, interpreting the output has hitherto been

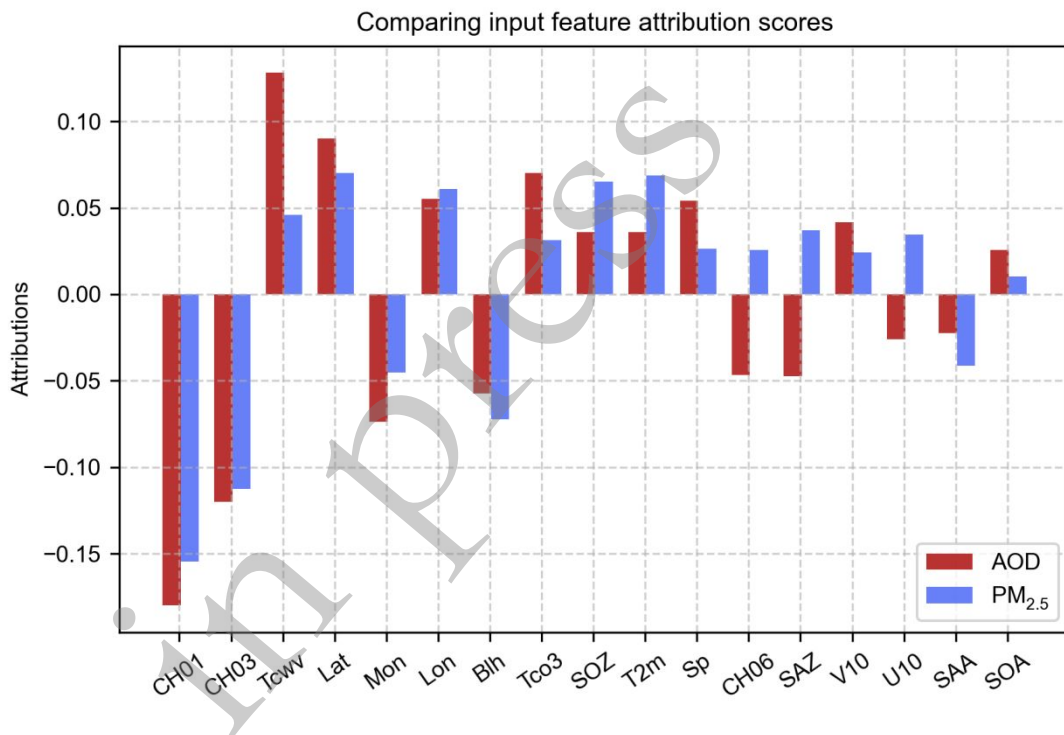
511 a demanding task. In this study, this issue is addressed by employing the DeepLift

512 algorithm from the Captum core library. The DeepLift algorithm allows us to

513 compare model predictions to a reference, enabling the quantification of the  
514 importance of each feature through a modified backpropagation technique. It is worth  
515 noting that we set 0 as the reference baseline for all features when computing  
516 imputation values. Comprehensive Captum tutorials are available on the official  
517 website (<https://captum.ai/>).

518 Figure 8 shows the attribution score exhibiting the significance of the  
519 independent variable on the dependent variable (AOD or  $PM_{2.5}$ ). Positive attribution  
520 scores indicate that the feature positively contributes to the predicted value, while  
521 negative scores suggest the opposite. In general, the feature importance of the two  
522 tasks has a certain similarity owing to inner connection between two related tasks and  
523 the parameter sharing mechanism. Specifically, the attribution scores of the visible  
524 channels (CH01 and CH03) are relatively high and have negative directions,  $-0.19$   
525 and  $-0.15$  respectively. Larger values of these two features result in smaller values of  
526 corresponding estimation. When the reflectance of the visible channels is lower, it  
527 usually indicates a stronger aerosol extinction capability in the atmosphere. The  
528 attribution scores of Tc<sub>wv</sub> in AOD (0.12) and  $PM_{2.5}$  (0.04) are both positive values. In  
529 general, when Tc<sub>wv</sub> increases, it will cause more extinction by water vapor and also  
530 affect the aggregation and sedimentation of particulate matter, thereby increasing  
531 AOD and  $PM_{2.5}$  concentration. The relative lower attribution score of Tc<sub>wv</sub> in  $PM_{2.5}$   
532 may be associated with the fact that  $PM_{2.5}$  represents the dry mass concentration of  
533 fine particulate, which is hardly affected by water vapor.

534 Furthermore, the seasonal effects (Mon) and geographical factor (Lon and Lat)  
 535 are critical in AOD and PM<sub>2.5</sub> estimation, which represent different temporal and  
 536 spatial heterogeneity. It is worth noting that due to the independence of the two tasks,  
 537 there will also be some features with opposite attribution score signs (i.e., CH06, U10  
 538 and SAZ), which also indicates that the MTL model will have a certain negative  
 539 transfer due to different task requirements. Therefore, the MTL model requires high  
 540 correlation between different tasks.



541

542 **Fig. 8.** Attribution scores for MTL model input features. The red and blue bars  
 543 represent the importance of the features for the AOD and PM<sub>2.5</sub> estimation,  
 544 respectively. Positive values indicate a positive contribution to the estimation, while  
 545 negative values signify the opposite.



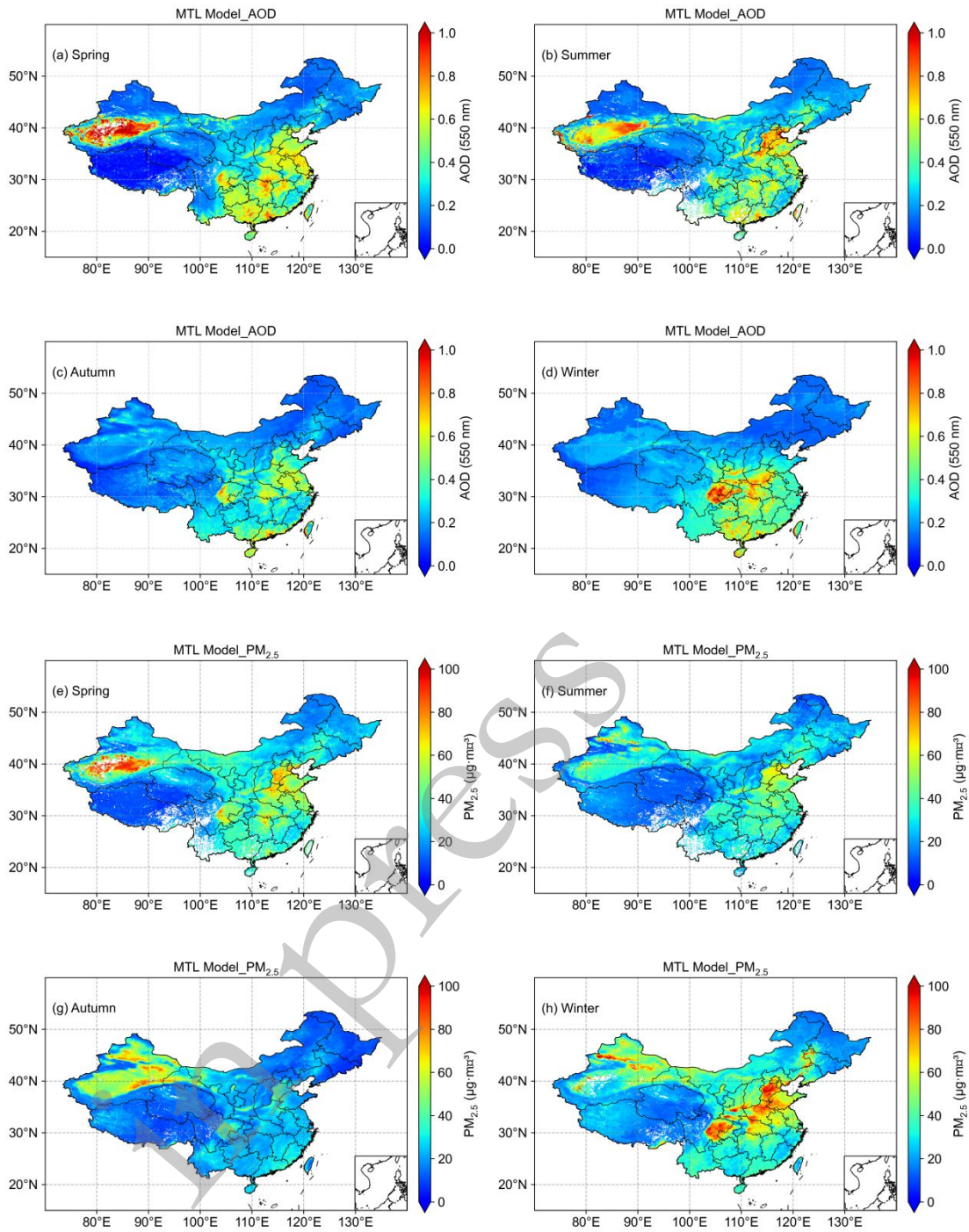
#### 546 *4.4 Spatial distribution of seasonal averages*

547 Figure 9 shows the spatial distribution of seasonal means of MTL AOD and  
548 MTL PM<sub>2.5</sub> for the period from March 12, 2018 to March 11, 2019. The seasonal  
549 mean values of MTL AOD are higher in spring (0.45) and summer (0.40) and relative  
550 lower in autumn (0.33) and winter (0.38), which is consistent with previous findings  
551 (Chen et al., 2023). The rise in aerosol loading in spring (Fig. 9a) can be attributed to  
552 frequent spring dust events in the north, leading to peaks in natural dust and  
553 windborne sand at the surface (He et al., 2016). Higher Blh in summer lead to vertical  
554 transport of aerosol particles to higher altitudes, which further enhances complete  
555 mixing of aerosols with water within the boundary layer. As a result, smaller aerosol  
556 particles within the boundary layer grow to optically active sizes (Qu et al., 2016).  
557 Higher temperatures in summer enhance photochemical reactions and also lead to an  
558 increase in aerosol loading during this season (Qi et al., 2013). Autumn is usually  
559 accompanied by stable atmospheric circulation and favorable diffusion conditions that  
560 favor dispersion and dilution of particulate matter, thus reducing aerosol loading. In  
561 contrast, in winter, lower Blh may result in particulate matter not being efficiently  
562 transported and mixed to higher altitudes, leading to a reduction in AOD (Qu et  
563 al., 2016). However, higher spatial distributions of AOD may still occur in areas with  
564 high emissions and complex topography, such as Sichuan and Chongqing, where  
565 special topography makes aerosol transport difficult.

566 Previous studies have indicated that PM<sub>2.5</sub> concentrations are higher in winter

567 and lower in summer (Li et al., 2017; Leung et al., 2020), which is similar to the MTL  
568 results.  $PM_{2.5}$  concentration are generally low in summer ( $26.45 \mu\text{g}\cdot\text{m}^{-3}$ ) and autumn  
569 ( $30.81 \mu\text{g}\cdot\text{m}^{-3}$ ) in China and showed similar spatial distributions (Fig. 9f and Fig. 9g).  
570 On the contrary,  $PM_{2.5}$  concentrations were significantly higher in spring ( $37.67$   
571  $\mu\text{g}\cdot\text{m}^{-3}$ ) and winter ( $41.51 \mu\text{g}\cdot\text{m}^{-3}$ ), especially in winter (Fig. 9e and Fig. 9h). The  
572 main reasons are the frequent sandstorms and the long-distance transmission of sand  
573 and dust in spring and the burning of coal and fossil fuels for heating in winter leading  
574 to more pollutant emissions in northern China (Wei et al., 2021).

575 A comparison of seasonal AOD and  $PM_{2.5}$  distribution reveals elevated aerosol  
576 levels over the North China Plain (NCP), Sichuan Basin, and Chongqing, attributed to  
577 factors such as intense human activities, adverse climatic conditions, and geographical  
578 features like basins that intensify anthropogenic aerosol emissions (Wang et al.,  
579 2018). Conversely, Northeast China, including Heilongjiang and Jilin, as well as  
580 Southwest China, such as Tibet and Yunnan, exhibit comparatively lower AOD and  
581  $PM_{2.5}$  levels due to sparse populations and reduced anthropogenic aerosol emissions.  
582 Furthermore, the favorable terrain and climatic conditions in these regions facilitate  
583 the dispersion of pollutants (Su et al., 2018).

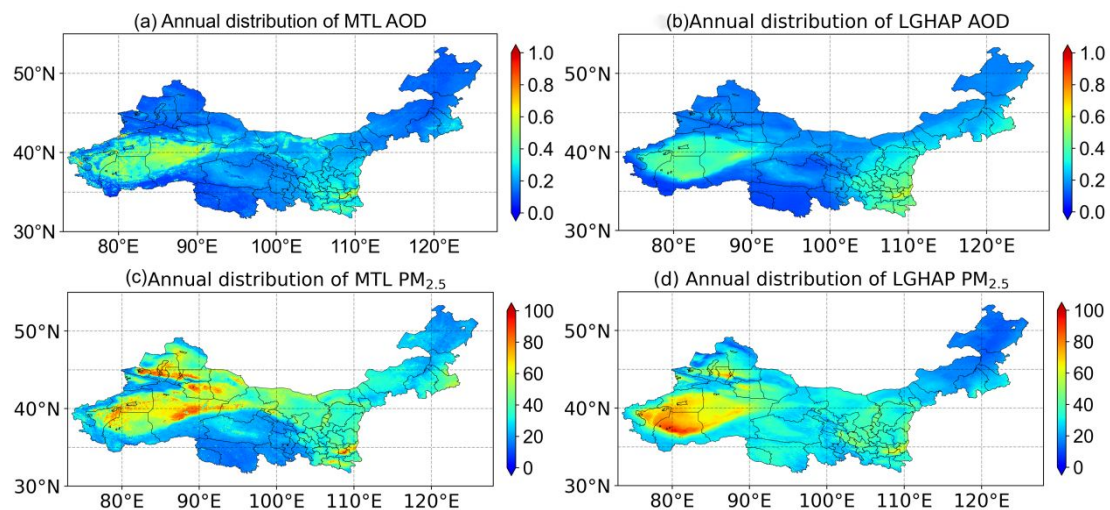


584

585 **Fig. 9.** Spatial distribution of seasonal mean AOD and  $PM_{2.5}$  in China in 2018 for (a),  
 586 (e) spring (March to May); (b), (f) summer (June to August); (c), (g) autumn  
 587 (September to November); and (d), (h) winter (December to February).

588

589 Fig. 10a and Fig. 10b show the spatial distribution of MTL AOD and LGHAP  
590 AOD, which have high consistency. The high values of MTL AOD and LGHAP  
591 AOD are located in the Taklamakan Desert and Guanzhong region of Shaanxi, and  
592 the annual average AOD ranges from 0.35 to 0.70. In Qinghai, northern Xinjiang, and  
593 the vast portion of Inner Mongolia, the annual average values of AOD are generally  
594 less than 0.3. Fig. 10c and Fig. 10d show spatial distribution of MTL PM<sub>2.5</sub> and  
595 LGHAP PM<sub>2.5</sub>, with annual mean PM<sub>2.5</sub> values exceeding 50  $\mu\text{g}\cdot\text{m}^{-3}$  in the Tarim  
596 Basin, northern Xinjiang, and Guanzhong region of Shaanxi. The topographic and  
597 climatic characteristics of the Tarim Basin restrict air movement and dispersion,  
598 leading to the accumulation of pollutants in the basin. The Guanzhong region of  
599 Shaanxi has well-developed industrial and transportation activities, resulting in a large  
600 amount of near-surface release of particulate matter. Overall, the annual mean value  
601 of LGHAP PM<sub>2.5</sub> is significantly higher than that of MTL PM<sub>2.5</sub>, especially in Qinghai  
602 Province, where LGHAP PM<sub>2.5</sub> is about 30  $\mu\text{g}\cdot\text{m}^{-3}$  compared to MTL PM<sub>2.5</sub> of about  
603 20  $\mu\text{g}\cdot\text{m}^{-3}$ . The main reason is that the LGHAP dataset is a gap-free aerosol product  
604 generated by reconstructing MODIS daily AOD gaps, while the MTL model needs to  
605 retrieve AOD under cloud-free conditions. The presence of clouds reduces the spatial  
606 and temporal continuity of the MTL model estimates, which results in the differences  
607 in the spatial distributions of AOD and PM<sub>2.5</sub>.



608

609 **Fig. 10.** In the figure, (a) and (b) are the spatial distribution maps of MTL AOD and  
 610 LGHAP AOD in Northwest China, and (c) and (d) are the spatial distribution maps of  
 611 MTL PM<sub>2.5</sub> and LGHAP PM<sub>2.5</sub> in Northwest China.

## 612 5 Conclusion

613 A high-precision and high-efficiency method for aerosol observation is of great  
 614 significance for ameliorating air quality, mitigating climate change, protecting  
 615 ecosystem, and improving human health. In order to improve the retrieval accuracy  
 616 and efficiency of AOD and PM<sub>2.5</sub>, a MTL algorithm based on parameter sharing  
 617 mechanism is proposed in this study. The constructed MTL model is able to retrieve  
 618 AOD and PM<sub>2.5</sub> simultaneously based on FY-4A AGRI data and ERA5 reanalysis.  
 619 The main conclusions are as follows.

620 (1) The estimated AOD and PM<sub>2.5</sub> from three models were evaluated at the  
 621 testing sites. The results showed that the MTL model predicted AOD ( $R^2 = 0.88$ ,  
 622 RMSE = 0.10, Within-EE = 85.70%) and PM<sub>2.5</sub> ( $R^2 = 0.84$ , RMSE = 13.76  $\mu\text{g}\cdot\text{m}^{-3}$ ,  
 623 Within-EE = 56.12%) the best compared to the conventional STL models (RF and

624 DNN). Furthermore, independent validation against AERONET sites in China and  
625  $PM_{2.5}$  observation sites in HK confirmed the generalization and reliability of MTL  
626 model (AOD:  $R^2 = 0.79$ , RMSE = 0.14, Within-EE = 74.09% and  $PM_{2.5}$ :  $R^2 = 0.76$ ,  
627 RMSE =  $8.11 \mu\text{g}\cdot\text{m}^{-3}$ , Within-EE = 61.35%). The performance of MTL model in  
628 Northwest China is mainly affected by geographical conditions and the number of  
629 training data.

630 (2) The feature contribution score of MTL model is calculated by using  
631 attribution algorithm. The results show that the feature contribution distributions of  
632 the two tasks are similar, but a few features have opposite contributions to the two  
633 tasks, which is related to the independence of the tasks. Poor correlation between  
634 tasks can lead to negative transfer effects in multiple MTL models across different  
635 tasks.

636 (3) The MTL model was used to estimate the AOD and  $PM_{2.5}$  concentrations in  
637 China in 2018, and the spatial distribution map was drawn. The results showed that  
638 the highest and lowest AOD values were found in spring (0.45) and autumn (0.33). In  
639 contrast, the seasonal variation of  $PM_{2.5}$  is large, with the highest and lowest  $PM_{2.5}$   
640 concentrations in winter ( $41.51 \mu\text{g}\cdot\text{m}^{-3}$ ) and summer ( $26.45 \mu\text{g}\cdot\text{m}^{-3}$ ). In addition, the  
641 difference of spatial distribution in Northwest China mainly depends on the  
642 spatio-temporal continuity of MTL model estimation results, and an effective  
643 interpolation algorithm can improve the integrity of MTL model estimation results.

644 The results of this study are based on the sample data collected from 1590

645 ground-based sites in China. However, the distribution of these sites is uneven, which  
646 further affects the overall training effect of the model, therefore, the estimation results  
647 in Northwest China have further room for improvement. In future work, using more  
648 efficient cloud detection algorithms and interpolation algorithms can further improve  
649 the accuracy and completeness of MTL model estimation results.

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